

RESEARCH ARTICLE

Exposure of tropical ecosystems to artificial light at night: Brazil as a case study

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Abstract

Artificial nighttime lighting from streetlights and other sources has a broad range of biological effects. Understanding the spatial and temporal levels and patterns of this lighting is a key step in determining the severity of adverse effects on different ecosystems, vegetation, and habitat types. Few such analyses have been conducted, particularly for regions with high biodiversity, including the tropics. We used an intercalibrated version of the Defense Meteorological Satellite Program's Operational Linescan System (DMSP/OLS) images of stable nighttime lights to determine what proportion of original and current Brazilian vegetation types are experiencing measurable levels of artificial light and how this has changed in recent years. The percentage area affected by both detectable light and increases in brightness ranged between 0 and 35% for native vegetation types, and between 0 and 25% for current vegetation (i.e. including agriculture). The most heavily affected areas encompassed terrestrial coastal vegetation types (restingas and mangroves), Semideciduous Seasonal Forest, and Mixed Ombrophilous Forest. The existing small remnants of Lowland Deciduous and Semideciduous Seasonal Forests and of Campinarana had the lowest exposure levels to artificial light. Light pollution has not often been investigated in developing countries but our data show that it is an environmental concern.

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Data availability statement: A complete data set for the night time images from Defense Meteorological Satellite Program's Operational Linescan System (DMSP/OLS) can be downloaded at the website of National Oceanic and Atmospheric Administration/National Geophysical Data Center (NOAA/NGDC) Earth Observation Group (<http://ngdc.noaa.gov/eog/>). The shapefile of Brazilian vegetation types can be accessed through REDD-PAC website (http://www.redd-pac.org/new_page.php?contents=data.csv) in WFS (web feature service) format. The authors state that they did not

Introduction

The nighttime environment is undergoing a dramatic transformation across the Earth's surface. The cycles of natural light (daily, lunar and seasonal) that have been major forms of environmental variation since the first emergence of life are being disrupted through the introduction of artificial lighting. A diversity of sources (including street lighting, advertising lighting, architectural lighting, security lighting, domestic lighting and vehicle lighting) are causing direct illumination as well as via skyglow, the scattering by atmospheric molecules or aerosols of artificial light at night that is emitted or reflected upwards [1–5].

Because natural cycles of light have previously provided rather consistent resources and sources of information for organisms, artificial nighttime lighting has a broad range of biological effects [5–7]. These span from gene to ecosystem levels [8,9]. They include effects on the

produce these data and do not claim ownership of the data.

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physiology, behaviour, reproductive success and mortality of species (e.g. [10–13]), on their abundance and distribution [14], and in turn on community structures and functioning (e.g. [2,15]). Moreover, it seems likely that the impacts of artificial nighttime lighting interact with those of other pressures on biodiversity, including habitat loss, climate change, other forms of pollution, and invasive species [16].

Determining the severity of these biological impacts rests, in part, on understanding of the spatial and temporal levels and patterns of artificial nighttime lighting, and particularly how these interact with those of different ecosystem, vegetation and habitat types [16]. At a global scale, virtually all natural terrestrial ecosystem types experience some level of exposure to artificial nighttime lighting or skyglow, and those that have been most and least affected have been identified [4]. However, more detailed regional analyses have largely been wanting. A few evaluations exist of regional patterns of artificial nighttime lighting, but these have not tended to determine the interaction with ecosystem, vegetation, or habitat types (e.g. [15,17]). Of particular concern is that work on spatial patterns of artificial nighttime lighting has focussed predominantly on China, Europe and North America [1,3,17,18] with almost no attention to global biodiversity hotspots. In particular, the potential environmental impacts of artificial nighttime lighting in tropical regions have been surprisingly little considered.

Aside from the often much greater levels of biodiversity that could be influenced, it remains unknown whether artificial nighttime lighting has different impacts in tropical regions compared with temperate ones. Obvious differences between tropical and non-tropical regions that might be significant are the short and rather invariant tropical periods of twilight, relatively low proportions of crepuscular and cathemeral species in tropical regions [19], the greater specialisation in tropical regions of some interspecific interactions that are known to be susceptible to influences from artificial nighttime light (e.g. plant-pollinator; [20,21]), and the prevalence of terrestrial species using bioluminescence, which are known to be vulnerable to light pollution [22–24].

In this paper we determine the spatial and temporal patterns of artificial nighttime lighting across Brazil in relation to the distribution of vegetation types. Brazil makes a particularly valuable case study. As well as being the largest country in South America, it has the largest number of species of any country in the world for many major taxonomic groups [25], has high levels of species endemism, and two recognised global biodiversity hotspots [26]. Brazil also has the richest biodiversity of bioluminescent beetles in the world [27].

Methods

Light data

Following Bennie *et al.* [3], we used nighttime stable lights annual composite images, created with data from the Defense Meteorological Satellite Program's Operational Linescan System (DMSP/OLS), downloaded from the National Oceanic and Atmospheric Administration archives (1992–2012, $n = 21$). These images capture upwardly reflected and directed nighttime light. The images are nominally at 1 km resolution, but are re-sampled from data at an equal angle of approximately 2.7 km resolution at the equator. These images cover spectral responses from 440 to 940 nm with the highest sensitivity in the 500 to 650 nm region. The spectral range encompasses the primary emissions from the most widely used sources for external lighting in Brazil: low pressure sodium (589 nm), high pressure sodium (from 540 nm to 630 nm) and mercury vapour (545 and 575 nm) [1,28].

Each pixel is represented by a digital number (DN) of between 0 and 63. Zero represents no detectable upward radiance, while brightly lit areas saturate at values of 63. Images were inter-calibrated and drift-corrected following the method of Bennie *et al.* [3]. An average calibrated

image for both the first (1992–1996) and the last (2008–2012) five years was created and the difference was calculated. To assess the changes over the full period time, we considered pixels increasing or decreasing by more than a threshold of 3 DN units of difference between the averages of the first and last years. It was previously observed that over 94% of observed increases in DN of more than 3 units and over 93% of observed decreases of the same magnitude were consistently related to the directions of changes on the ground (e.g., expansion or contraction of urban and industrial areas) [3]. Following Gaston *et al.* [29] and Duffy *et al.* [30], we considered pixels as exposed to artificial light when they had values higher than 5.5 DN units. By using a threshold effectively twice the detection limit for change, we defined a conservative estimate of lit area and limited the extent to which dark sites may be classified as lit due to noise in the data set or calibration errors [29,30].

Vegetation type data

We used the vegetation map produced by the Brazilian Institute for Geography and Statistics [31], which is recommended as a good basis to compare with data obtained from remote sensing images [32]. This map presents both original native vegetation and current vegetation and land cover. The former portrays the original vegetation classes in Brazil likely found at the time of Portuguese colonisation [31], and the latter describes the vegetation now present [31]. Original vegetation includes 24 wider classes while the current is more detailed, including 52 classes (Table 1). The shapefile was produced by IBGE—Brazilian Institute of Geography and Statistics and accessed through REDD-PAC website (http://www.redd-pac.org/new_page.php?contents=data.csv) in WFS (web feature service) format.

The IBGE map divides vegetation into two broad classes: forests and non-forests [33]. Forests are divided into Ombrophilous Forest and Seasonal Forest. The former is further divided into three physiognomies (Dense, Open and Mixed) and the last into two (Deciduous and Semi-deciduous). All of these can be classified by up to five formations: Alluvial, Lowland, Submontane, Montane and High-montane (Table 1). Non-forests are divided into four formations: Campinarana, Savanna, Steppe-savanna, and Steppe, which in turn can be divided into up to four formations: Forest, Woody, Shrubland, and Grassland. The map also classifies pioneer formations—that encompass vegetation influenced by rivers (Alluvial Areas), by the sea (Restingas), and by both (Mangroves)—Ecotones, Relict Vegetation and Water. When considering the current vegetation, it also includes Agriculture and Secondary Vegetation classes (Table 1).

Processing

To define the proportional area of each vegetation type that has been exposed to artificial nighttime light, we overlaid both original and current vegetation shapefiles on the DMSP data for the most recent five years (2008–2012). We extracted both the number of lit pixels and the total number of pixels inside each vegetation type and divided the first by the second. To assess changes, we overlaid the two vegetation shapefiles on the difference between the first (1992–1996) and the last (2008–2012) five years of DMSP data. We extracted the number of increasing pixels, decreasing pixels and the total number of pixels inside each vegetation type. We divided the number of increasing and decreasing pixels by the total in each vegetation type, achieving the proportional area where artificial light has been increasing and decreasing respectively.

Results

Overall, the percentage of area of each vegetation type affected by increases in artificial light was higher than the percentages affected by ‘detectable’ light (Figs 1 and 2). Less than

Table 1. Vegetation classification for Brazil according to IBGE (2012).

Forest	Ombrophilous Forest	Dense Ombrophilous Forest	Alluvial Dense Ombrophilous Forest	
			Lowland Dense Ombrophilous Forest	
			Sub-Montane Dense Ombrophilous Forest	
			Montane Dense Ombrophilous Forest	
			Open Ombrophilous Forest	
			Alluvial Open Ombrophilous Forest	
	Seasonal Forest	Semi-deciduous Seasonal Forest	Lowland Open Ombrophilous Forest	
			Sub-Montane Open Ombrophilous Forest	
			Mixed Ombrophilous Forest	
			Montane Mixed Ombrophilous Forest	
			High-montane Mixed Ombrophilous Forest	
			Alluvial Semi deciduous Seasonal Forest	
Deciduous Seasonal Forest	Deciduous Seasonal Forest	Lowland Semi deciduous Seasonal Forest		
		Sub-Montane Semi-deciduous Seasonal Forest		
		Montane Semi-deciduous Seasonal Forest		
		Lowland Deciduous Seasonal Forest		
		Sub-Montane Deciduous Seasonal Forest		
		Montane Deciduous Seasonal Forest		
Non Forest	Campinarana	Campinarana	Forest Campinarana	
			Woody Campinarana	
			Shrubland Campinarana	
			Grassland Campinarana	
			Savanna	
			Forest Savanna	
			Woody Savanna	
			Parkland Savanna	
			Grassland Savanna	
			Steppe-savanna	
			Forest Steppe-savanna	
			Woody Steppe-savanna	
	Parkland Steppe-savanna			
	Grassland Steppe-savanna			
	Steppe	Steppe	Steppe	Woody Steppe
				Parkland Steppe
				Grassland Steppe
				Pioneer formation
				Alluvial Areas
				Restinga
	Other	Ecotone	Campinarana/Ombrophilous Forest	Mangrove
				Campinarana/Ombrophilous Forest
				Steppe/seasonal Forest
				Steppe/seasonal Forest
Seasonal Forest /Primary Formations				
Seasonal Forest /Primary Formations				
Dense Ombrophilous Forest/Mixed Ombrophilous Forest				
Dense Ombrophilous Forest/Mixed Ombrophilous Forest				
Ombrophilous Forest/Seasonal Forest				
Ombrophilous Forest/Seasonal Forest				
Steppe savanna /Seasonal Forest				
Steppe savanna /Seasonal Forest				
Savanna/Seasonal Forest				
Savanna/Seasonal Forest				
Savanna/Ombrophilous Forest				
Savanna/Ombrophilous Forest				
Savanna/Primary Formations				
Savanna/Primary Formations				
Savanna/Steppe-savanna				
Savanna/Steppe-savanna				
Savanna/Steppe-savanna/Seasonal Forest				
Savanna/Steppe-savanna/Seasonal Forest				
Relict Vegetation	Relict Vegetation	Relict Vegetation	High-montane Relict Vegetation	
			Montane Relict Vegetation	
Water	Water	Water	Coastal Water Mass	

(Continued)

Table 1. (Continued)

			Continental Water Mass
	Rocky Outcrops	Rocky Outcrops	Rocky Outcrops
			Agriculture
			Secondary Vegetation

The third column corresponds to original vegetation and the fourth column to current vegetation.

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0.00001% of the areas of vegetation types experienced decreases in brightness so we considered only the increases in the results.

Spatial distribution of detectable light and increases in brightness followed similar patterns. The most affected areas were strongly concentrated along the coast, in the east, particularly in the southeast, while less affected areas were located in the west and in the central region (Fig 3A and 3B).

Pre-colonization native vegetation

The area of original vegetation types affected by both detectable light and increases in brightness ranged between 0% and approximately 35%. Types affected by detectable light in more than 10% of their areas include pioneer formations (which encompass Mangroves, Restingas, and Alluvial Areas—Table 1), Semideciduous Seasonal Forest, Mixed Ombrophilous Forest, and six ecotones containing these ones and also Savanna, Steppe-savanna, Dense Ombrophilous Forest, and Steppe (Fig 1).

Less than 1% of the areas of three original vegetation types were affected by both detectable light and increases in exposure: Campinarana/Ombrophilous Forest, Savanna/ Pioneer Formations, and Ombrophilous Forest/Seasonal Forest (Fig 1). Two out of 24 original vegetation types had levels of detectable artificial light at night below the threshold: Rocky Outcrops and Campinarana (Fig 1). The less affected original vegetation types were concentrated in the west and in the central area while the most affected were in the southeast and northeast (Fig 3A and 3C).

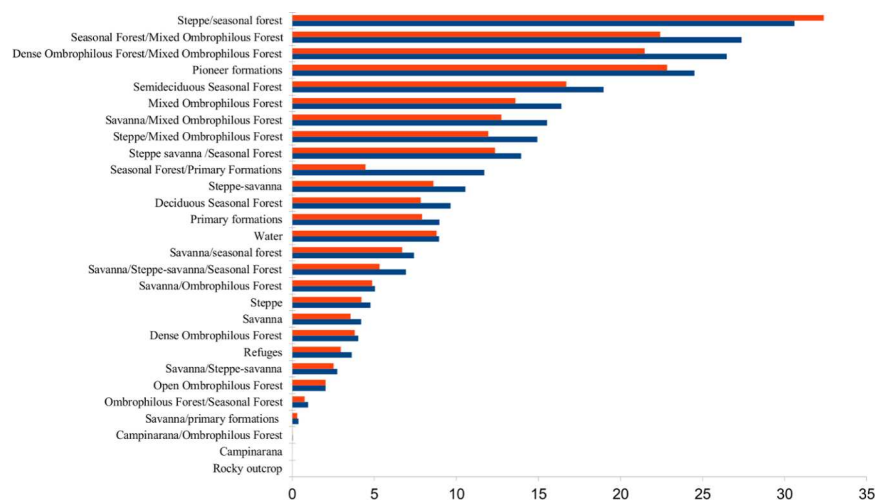


Fig 1. Percentage of area of original vegetation types affected by artificial light. Horizontal bars show the percentage of total land surface area occupied by each original vegetation type that had more than 5.5 Digital Number (DN) units in 2008–2012 (red) or an increase of more than 3 DN units between 1992–2012 and 2008–2012 (blue).

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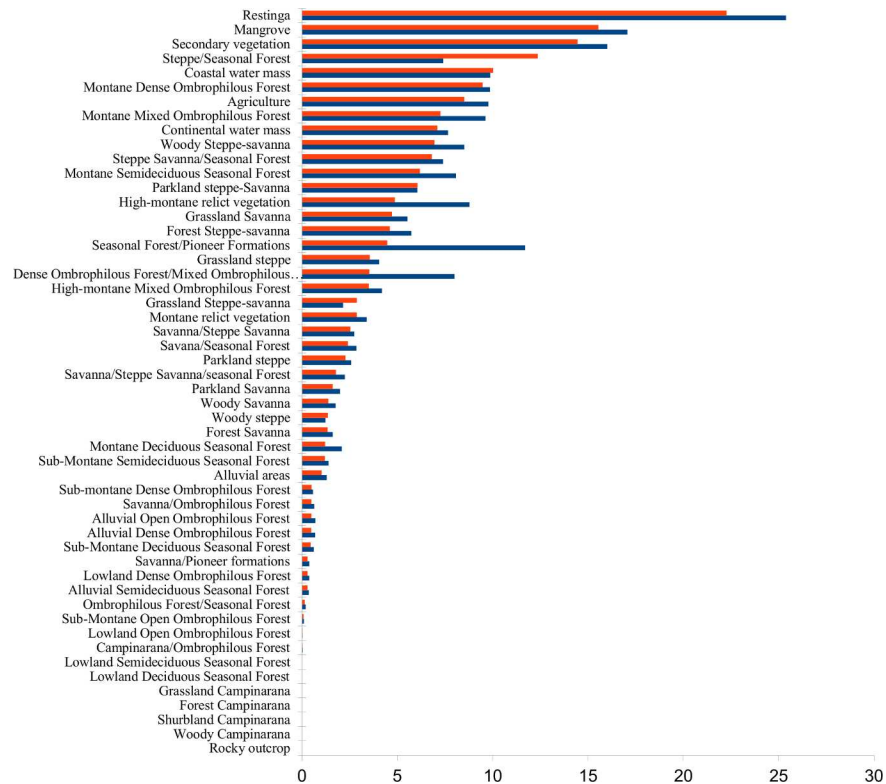


Fig 2. Percentage of area of current vegetation types affected by artificial light. Horizontal bars show the percentage of total land surface area occupied by each current vegetation type that had more than 5.5 Digital Number (DN) units in 2008–2012 (red) or an increase of more than 3 DN units between 1992–2012 and 2008–2012 (blue).

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Current vegetation

The area of current vegetation types affected by detectable light ranged between less than 1% and approximately 25%. Restingas, Mangroves, Secondary Vegetation, and Steppe/Seasonal Forest had more than 10% of their areas affected by detectable light (Fig 2). The first three were also the most affected by changes in brightness as well as Seasonal Forest/ Pioneer Formations (Fig 2).

Vegetation types with less than 1% of their areas affected by both detectable light and increases in exposure were the three formations of Open and Dense Ombrophilous Forest (Alluvial, Lowland and Sub-montane—Table 1), Alluvial Semideciduous Seasonal Forest, Sub-Montane Deciduous Seasonal Forest and four ecotones involving Savanna, Ombrophilous Forest, Pioneer Formations (mainly Mangroves and Restingas), Campinarana and Seasonal Forest (Fig 2).

100% of the areas of seven of the 52 current vegetation types had levels of detectable artificial light lower than the threshold: Rocky Outcrops, the four formations of Campinarana (i.e. Woody, Shrubland, Forest and Grassland—Table 1), Lowland Deciduous Seasonal Forest and Lowland Semideciduous Seasonal Forest (Fig 2).

The most affected current vegetation types were strongly concentrated along the coast, in the east. The less affected ones occurred in the west (where Amazonia rainforest is located) and in the central area (Fig 3B and 3D).

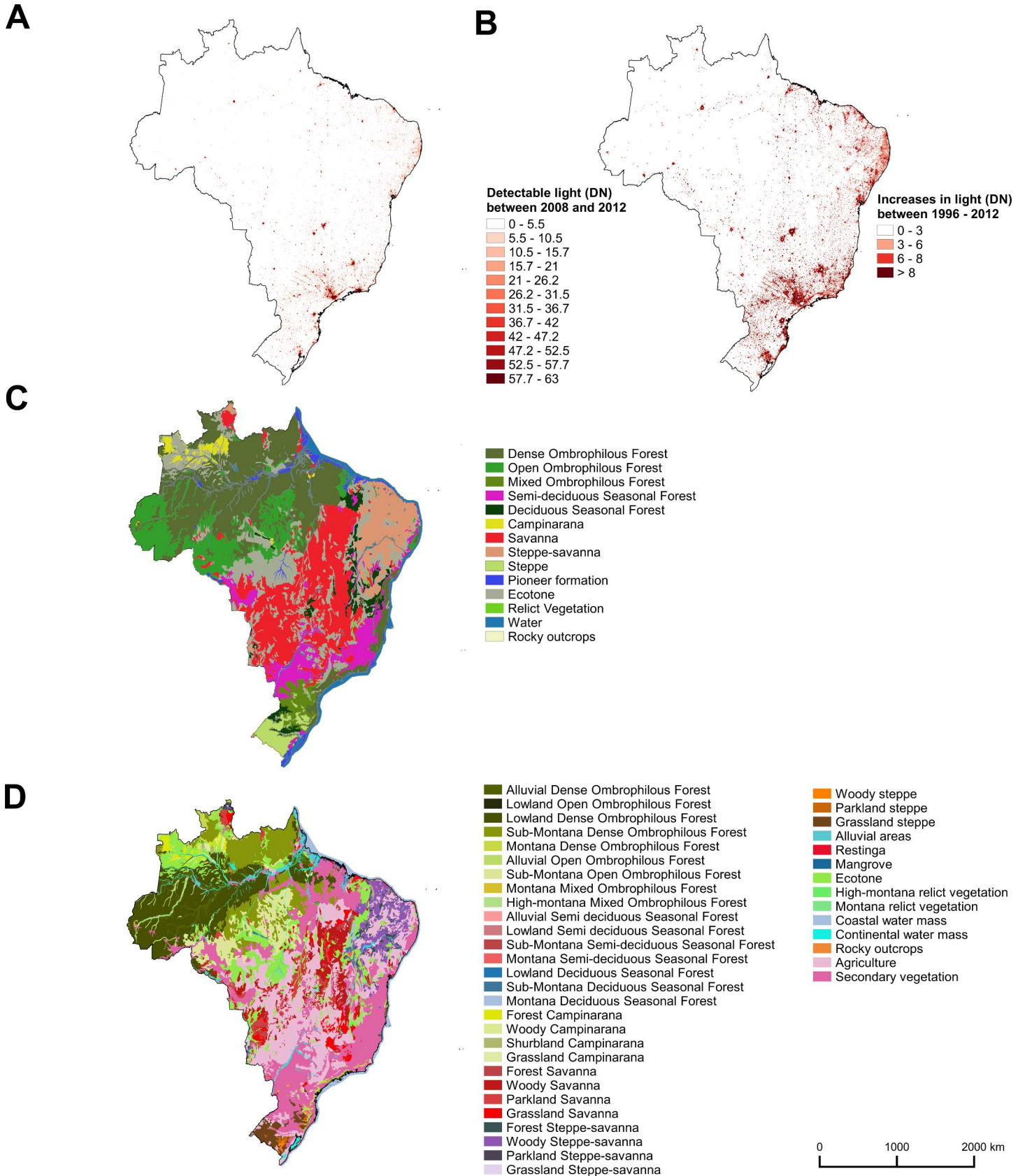


Fig 3. Spatial distribution of artificial light and vegetation types in Brazil. Distribution of: (A) pixels with detectable light (DN > 5.5) in the most recent five years (2008–2012); (B) pixels with increases in brightness (differences higher than 3 DN) between the first (1992–1998) and the last (2008–2012) five years; (C) original vegetation types; and (D) current vegetation types. The figure was created using QGIS 2.12.3. Nighttime light images were created with data from the Defense Meteorological Satellite Program's Operational Linescan System (DMSP/OLS), freely available at the website of National Oceanic and Atmospheric Administration/National Geophysical Data Center (NOAA/NGDC) Earth Observation Group (<http://ngdc.noaa.gov/eog/>). The shapefile of Brazilian vegetation types was produced by IBGE (Brazilian Institute of Geography and Statistics) and is freely available at REDD-PAC website (http://www.redd-pac.org/new_page.php?contents=data.csv) in WFS (web feature service) format.

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Discussion

In this paper we provide the first assessment of the broad level of exposure of tropical and sub-tropical ecosystems to artificial light at night at a regional extent. Because the percentage of areas of the different vegetation types affected by increases in brightness was higher than those affected by detectable light in most of the cases (Figs 1, 2, 3A and 3B), it seems inevitable that the extent of artificial lighting will continue to increase.

The highest aggregations of artificial lights in Brazil are in the coastal regions (Fig 3A and 3B) from where occupancy of Brazilian territory by Europeans started and where the larger urban agglomerations are now located [34]. The three most widely lit vegetation types when considering original vegetation are ecotones and all of them involve Seasonal Forest or Mixed Ombrophilous Forest (Fig 1). Semideciduous Seasonal Forest and Mixed Ombrophilous Forest themselves are also widely lit by detectable light (16.7% and 13.6% respectively—Fig 1). These levels of coverage by artificial lighting are lower for current vegetation of the same types (6.18% for Montane Semideciduous Forest, 1.2% for Sub-montane Semideciduous Forest, 7.25% for Montane Mixed Ombrophilous Forest, and 3.5% for High-montane Mixed Ombrophilous Forest) because they have been highly converted and the current remnants are small [35]. Of the current vegetation types, Restingas, Mangroves and Coastal water mass are among the five with the greatest percentage coverage by artificial nighttime lighting (Fig 2).

Imagery of emissions of upward radiance are the best available data to assess both the presence and trends in artificial light at a regional scale (other artificial nighttime lighting data sets do not yet capture trends). However, as pointed by Bennie *et al.* [4], trends established using these data must be interpreted with caution because the relationships between the images captured by the satellites and biologically relevant levels of light experienced by species are not straightforward. First, the spectral response of the OLS instrument covers the ranges of the most commonly used sources for external light, which differs from the action spectra of biological processes depending on the species. Second, because DMSP/OLS images are approximately at 2.7 km resolution, the correspondence between the illuminated areas in the images and the areas at the ground surface where biologically significant levels of lights are present is not precise. And finally, upwards radiance measures do not encompass horizontal emissions or skyglow—although it is important to observe that empirical data on temporal trends in the spatial occurrence of skyglow at continental scales are not presently available, and modelled surface data have large uncertainties [36,37].

Whilst an impressively wide array of ecological impacts of artificial nighttime lighting have been documented (see Introduction), the most important effects on given vegetation types and their associated communities remain unknown. Nonetheless, Semideciduous Seasonal Forest may potentially be differentially impacted because the trees lose from 20% to 50% of their leaves during the unfavourable season (i.e. dry and cold season in tropical and subtropical zones respectively [30]) and street lighting has previously been shown in other contexts to affect leaf fall timing as well as the speed of leaf growth [38,39]. Mixed Ombrophilous Forest, also known as araucaria forest due to the dominance of Brazilian pine (*Araucaria angustifolia*) [33], has a notably high richness and diversity of dung beetles [40]. It is known that dung

beetles exploit moonlight, the celestial polarization pattern and the starry sky for orientation [41–44]. Given the important role of dung beetles in decomposition and nutrient cycling in tropical ecosystems, it seems likely that the high levels of artificial light and increase in brightness found in Ombrophilous Mixed Forest will affect its functioning.

Both Restinga and Mangrove are heavily overlapped by artificial light. Restinga is the terrestrial pioneer vegetation that occurs on sandy shore environments, especially on dunes, and is directly influenced by the sea [33]. Restinga harbours a high diversity of bats [45–47], which are known to be important for the maintenance of forests and to be disturbed by artificial light [48–50]. Around the world, mangroves are threatened by deforestation, illegal shrimp culture, expansion of urban areas, tourism, fishing and pollution [51]. Nine percent of the global area of natural or semi natural mangroves has seen an increase in exposure to artificial light [4]. In Brazil this percentage is 17% in the same period (Fig 2), with more than 15% of the mangrove area experiencing detectable light (Fig 2). Given that Brazil accounts for approximately 50% of mangroves in South America and 7% of the world's mangroves [51], light pollution in these areas should be of particular concern. Both Restinga and Mangrove are coastal ecosystems and the coastal water mass itself is also highly affected by light (Fig 2). Five out of seven extant species of marine turtles in the world nest on the Brazilian coast (*Chelonia mydas*, *Caretta caretta*, *Dermochelys coriacea*, *Eretmochelys imbricata*, and *Lepidochelys olivacea*)—all of them are listed as threatened on the IUCN Red List (<http://www.iucnredlist.org/search>). Artificial lighting disrupts sea turtle hatchling orientation from the nest to the sea [52]. To protect Brazilian coastal ecosystems, the law forbids illumination within 50 m of the beach strip between Rio de Janeiro and Rio Grande do Norte States—which corresponds to approximately 2 500 km out of the 7 367 km of Brazilian coast [53]. Due to the scarcity of studies on the consequence of light pollution in these ecosystems, it is not possible to assess if the law is effective.

In most developing countries artificial nighttime lighting is relatively recent and concentrated in dense populated urban areas [37]. In contrast, in highly industrialised countries it is much more widespread [1,4], and often considered thus to be a much greater concern. However, our results here highlight that lighting is extensive in some developing countries, including ones with exceptionally high levels of biodiversity. These results also suggest that it is still possible to find vegetation types with natural sky background brightness. Countries in which this is the case have the opportunity to base policies, regulations, and guidelines on minimising rather than mitigating the ecological impacts of artificial nighttime lighting.

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Conceptualization: JRF KJG.

Data curation: JRF.

Formal analysis: JRF.

Funding acquisition: JRF WM.

Investigation: JRF.

Methodology: JB.

Resources: KJG.

Supervision: KJG.

Writing – original draft: JRF.

Writing – review & editing: WM JB KJG.

References

1. Cinzano P, Falchi F, Elvidge CD. The first World Atlas of the artificial night sky brightness. *Mon Not R Astron Soc.* 2001; 328: 689–707.
2. Davies TW, Bennie J, Inger R, Gaston KJ. Artificial light alters natural regimes of night-time sky brightness. *Sci Rep.* 2013; 3: 1722.
3. Bennie J, Davies TW, Duffy JP, Inger R, Gaston KJ. Contrasting trends in light pollution across Europe based on satellite observed night time lights. *Sci Rep.* 2014; 4: 1–9.
4. Bennie J, Duffy J, Davies T, Correa-Cano M, Gaston K. Global trends in exposure to light pollution in natural terrestrial ecosystems. *Remote Sens.* 2015; 7: 2715–2730.
5. Gaston KJ, Bennie J, Davies TW, Hopkins J. The ecological impacts of nighttime light pollution: A mechanistic appraisal. *Biol Rev.* 2013; 88: 912–927. doi: [10.1111/brv.12036](https://doi.org/10.1111/brv.12036) PMID: [23565807](https://pubmed.ncbi.nlm.nih.gov/23565807/)
6. Kyba CCM, Hötker F. Do artificially illuminated skies affect biodiversity in nocturnal landscapes? *Landsc Ecol.* 2013; 28: 1637–1640.
7. Hötker F, Wolter C, Perkin EK, Tockner K. Light pollution as a biodiversity threat. *Trends Ecol Evol.* 2010; 25: 681–682. doi: [10.1016/j.tree.2010.09.007](https://doi.org/10.1016/j.tree.2010.09.007) PMID: [21035893](https://pubmed.ncbi.nlm.nih.gov/21035893/)
8. Ashkenazi L, Haim A. Light interference as a possible stressor altering HSP70 and its gene expression levels in brain and hepatic tissues of golden spiny mice. *J Exp Biol.* 2012; 215: 4034–4040. doi: [10.1242/jeb.073429](https://doi.org/10.1242/jeb.073429) PMID: [22933613](https://pubmed.ncbi.nlm.nih.gov/22933613/)
9. Gaston KJ, Visser ME, Hötker F. The biological impacts of artificial light at night: the research challenge. *Phil Trans R Soc B.* 2015; 370: 20140133. doi: [10.1098/rstb.2014.0133](https://doi.org/10.1098/rstb.2014.0133) PMID: [25780244](https://pubmed.ncbi.nlm.nih.gov/25780244/)
10. Arendt J. Melatonin and the pineal gland: influence on mammalian seasonal and circadian physiology. *Rev Reprod.* 1998; 3: 13–22. PMID: [9509985](https://pubmed.ncbi.nlm.nih.gov/9509985/)
11. Le Tallec T, Perret M, Théry M. Light pollution modifies the expression of daily rhythms and behavior patterns in a nocturnal primate. *PLoS One.* 2013; 8.
12. van Geffen KG, van Grunsven RHA, van Ruijven J, Berendse F, Veenendaal EM. Artificial light at night causes diapause inhibition and sex-specific life history changes in a moth. *Ecol Evol.* 2014; 4: 2082–2089. doi: [10.1002/ece3.1090](https://doi.org/10.1002/ece3.1090) PMID: [25360250](https://pubmed.ncbi.nlm.nih.gov/25360250/)
13. Rodríguez A, Rodríguez B, Curbelo AJ, Pérez A, Marrero S, Negro JJ. Factors affecting mortality of shearwaters stranded by light pollution. *Anim Conserv.* 2012; 15: 519–526.
14. Gaston KJ, Bennie J. Demographic effects of artificial nighttime lighting on animal populations. *Environ Rev.* 2014; 8: 1–8.
15. Bennie J, Davies TW, Cruse D, Inger R, Gaston KJ, Gaston KJ. Cascading effects of artificial light at night: resource-mediated control of herbivores in a grassland ecosystem. *Philos Trans R Soc B.* 2015; 370: 20140131.
16. Gaston KJ, Duffy JP, Gaston S, Bennie J, Davies TW. Human alteration of natural light cycles: causes and ecological consequences. *Oecologia.* 2014; 176: 917–931. doi: [10.1007/s00442-014-3088-2](https://doi.org/10.1007/s00442-014-3088-2) PMID: [25239105](https://pubmed.ncbi.nlm.nih.gov/25239105/)
17. Miller MW. Apparent effects of light pollution on singing behavior of american robins. *Condor.* 2006; 108: 130–139.
18. Ma T, Zhou C, Pei T, Haynie S, Fan J. Quantitative estimation of urbanization dynamics using time series of DMSP/OLS nighttime light data: A comparative case study from China's cities. *Remote Sens Environ.* Elsevier Inc.; 2012; 124: 99–107.
19. Bennie JJ, Duffy JP, Inger R, Gaston KJ. Biogeography of time partitioning in mammals. *Proc Natl Acad Sci U S A.* 2014; 111: 13727–32. doi: [10.1073/pnas.1216063110](https://doi.org/10.1073/pnas.1216063110) PMID: [25225371](https://pubmed.ncbi.nlm.nih.gov/25225371/)
20. van Langevelde F, Ettema JA, Donners M, WallisDeVries MF, Groenendijk D. Effect of spectral composition of artificial light on the attraction of moths. *Biol Conserv.* Elsevier Ltd; 2011; 144: 2274–2281.
21. MacGregor CJ, Pocock MJO, Fox R, Evans DM. Pollination by nocturnal Lepidoptera, and the effects of light pollution: a review. *Ecol Entomol.* 2015; 40: 187–198. doi: [10.1111/een.12174](https://doi.org/10.1111/een.12174) PMID: [25914438](https://pubmed.ncbi.nlm.nih.gov/25914438/)
22. Viviani VR, Rocha MY, Hagen O. Fauna de besouros bioluminescentes (Coleoptera: Elateroidea: Lampyridae; Phengodidae, Elateridae) nos municípios de Campinas, Sorocaba-Votorantim e Rio Claro-Limeira (SP, Brasil): biodiversidade e influência da urbanização. *Biota Neotrop.* 2010; 10: 0–0.

23. Picchi MS, Avolio L, Azzani L, Brombin O, Camerini G. Fireflies and land use in an urban landscape: The case of *Luciola italica* L. (Coleoptera: Lampyridae) in the city of Turin. *J Insect Conserv*. 2013; 17: 797–805.
24. Hagen O, Santos RM, Schlindwein MN, Viviani VR. Artificial night lighting reduces firefly (Coleoptera: Lampyridae) occurrence in Sorocaba, Brazil. *Adv Entomol*. 2015; 24–32.
25. Lewinsohn TM, Prado PI. How many species are there in Brazil? *Conserv Biol*. 2005; 19: 619–624.
26. Myers N, Mittermeier RA, Mittermeier CG, da Fonseca GB, Kent J. Biodiversity hotspots for conservation priorities. *Nature*. 2000; 403: 853–858. doi: [10.1038/35002501](https://doi.org/10.1038/35002501) PMID: [10706275](https://pubmed.ncbi.nlm.nih.gov/10706275/)
27. Costa C. Estado de conocimiento de los Coleoptera neotropicales. Proyecto Iberoamericano de Biogeografía y Entomología Sistemática: PRIBES 2000: trabajos del 1er taller iberoamericano de entomología sistemática. SEA: Sociedad Entomológica Aragonesa; 2000. pp. 99–114.
28. World Bank Group. Iluminando Cidades Brasileiras—Modelos de negócio para Eficiência Energética em Iluminação Pública. 2016.
29. Gaston KJ, Duffy JP, Bennie J. Quantifying the erosion of natural darkness in the global protected area system. *Conserv Biol*. 2015; 29: 1132–1141. doi: [10.1111/cobi.12462](https://doi.org/10.1111/cobi.12462) PMID: [25693660](https://pubmed.ncbi.nlm.nih.gov/25693660/)
30. Duffy JP, Bennie J, Durán AP, Gaston KJ. Mammalian ranges are experiencing erosion of natural darkness. *Sci Rep*. 2015; 5: 12042. doi: [10.1038/srep12042](https://doi.org/10.1038/srep12042) PMID: [26155917](https://pubmed.ncbi.nlm.nih.gov/26155917/)
31. IBGE Instituto Brasileiro de Geografia e Estatística. Mapa de Vegetação do Brasil. Rio de Janeiro; 2004.
32. Buurman M, Câmara G, Carvalho AY de, Jones J, Cartaxo R, Mosnier A, et al. Description of the GLOBIOM-BRAZIL database available in the REDD-PAC WFS server. REDD-PAC REDD+ Policy Assessment Centre; 2015. pp. 1–68.
33. IBGE Instituto Brasileiro de Geografia e Estatística. Manual Técnico da Vegetação Brasileira. Série Manuais Técnicos em Geociências 1. 2a revisad. Rio de Janeiro; 2012.
34. IBGE Instituto Brasileiro de Geografia e Estatística. Atlas do Censo Demográfico 2010. Rio de Janeiro; 2013.
35. Morellato P, Haddad CFB. Introduction: The Brazilian Atlantic Forest. *Biotropica*. 2000; 32: 786–792.
36. Kyba CCM, Tong KP, Bennie J, Birriel I, Birriel JJ, Cool A, et al. Worldwide variations in artificial sky-glow. *Sci Rep*. 2015; 5: 8409. doi: [10.1038/srep08409](https://doi.org/10.1038/srep08409) PMID: [25673335](https://pubmed.ncbi.nlm.nih.gov/25673335/)
37. Falchi F, Cinzano P, Duriscoe D, Kyba CCM, Elvidge CD, Baugh K, et al. The new world atlas of artificial night sky brightness. 2016; 1–26.
38. Matzke EB. The effect of street lights in delaying leaf-fall in certain trees. *Am J Bot*. 1936; 23: 446.
39. Han B, Kim J, Kwak J, Choi T. Correlation between the illuminance and the flowering and leaf growth of trees at night—in case of downtown from Jamsil Station to Olympic Park, Seoul. 2015; 29: 441–453.
40. Campos RC, Hernández MIM. Dung beetle assemblages (Coleoptera, Scarabaeinae) in Atlantic forest fragments in southern Brazil. *Rev Bras Entomol*. 2013; 57: 47–54.
41. Byrne M, Dacke M, Nordström P, Scholtz C, Warrant E. Visual cues used by ball-rolling dung beetles for orientation. *J Comp Physiol A Neuroethol Sensory, Neural, Behav Physiol*. 2003; 189: 411–418.
42. Dacke M, Nilsson D-E, Scholtz CH, Byrne M, Warrant EJ. Insect orientation to polarized moonlight: An African dung beetle uses the moonlit sky to make a swift exit after finding food. *Nature*. 2003; 424: 33.
43. Dacke M, Byrne MJ, Scholtz CH, Warrant EJ. Lunar orientation in a beetle. *Proc Biol Sci*. 2004; 271: 361–5. doi: [10.1098/rspb.2003.2594](https://doi.org/10.1098/rspb.2003.2594) PMID: [15101694](https://pubmed.ncbi.nlm.nih.gov/15101694/)
44. Dacke M, Baird E, Byrne M, Scholtz CH, Warrant EJ. Dung beetles use the milky way for orientation. *Curr Biol*. 2013; 298–300. doi: [10.1016/j.cub.2012.12.034](https://doi.org/10.1016/j.cub.2012.12.034) PMID: [23352694](https://pubmed.ncbi.nlm.nih.gov/23352694/)
45. Lins J, Moraes L, Captivo E, Costa LA, Esbérard CEL. Bats from the Restinga of Praia das Neves, state of Espírito Santo, Southeastern Brazil. *Check List*. 2009; 5: 364–369.
46. Luz JL, Mangolin R, Esbérard CEL, Bergallo HDG. Morcegos (Chiroptera) capturados em lagoas do Parque Nacional da Restinga de Jurubatiba, Rio de Janeiro, Brasil. *Biota Neotrop*. 2011; 11: 161–168.
47. Oprea M, Esbérard CEL, Vieira TB, Mendes P, Pimenta VT, Brito D, et al. Bat community species richness and composition in a restinga protected area in Southeastern Brazil. *Braz J Biol*. 2009; 69: 1073–1079. PMID: [19967177](https://pubmed.ncbi.nlm.nih.gov/19967177/)
48. Stone EL, Jones G, Harris S. Street lighting disturbs commuting bats. *Curr Biol*. Elsevier Ltd; 2009; 19: 1123–1127. doi: [10.1016/j.cub.2009.05.058](https://doi.org/10.1016/j.cub.2009.05.058) PMID: [19540116](https://pubmed.ncbi.nlm.nih.gov/19540116/)
49. Lewanzik D, Voigt CC. Artificial light puts ecosystem services of frugivorous bats at risk. *J Appl Ecol*. 2014; 51: 388–394.

50. Day J, Baker J, Schofield H, Mathews F, Gaston KJ. Part-night lighting: Implications for bat conservation. *Anim Conserv*. 2015; 18: 512–516.
51. FAO—Food and Agriculture Organization of the United Nations. The world's mangroves 1980–2005: A thematic study prepared in the framework of the Global Forest Resources Assessment 2005. FAO Forestry Paper. Rome; 2007.
52. Lorne J, Salmon M. Effects of exposure to artificial lighting on orientation of hatchling sea turtles on the beach and in the ocean. *Endanger Species Res*. 2007; 3: 23–30.
53. IBAMA—Instituto Brasileiro do Meio Ambiente e dos Recursos Renováveis. Portaria No 11, De 30 De Janeiro De 1995. *Diário Oficial da União Brasil: Proíbe qualquer fonte de iluminação que ocasione intensidade luminosa superior a Zero Lux, numa faixa de praia entre linha de maior baixa-maré a 50m (cinquenta metros) acima da linha maior preamar do ano (maré de sizígia), entre os estados de Rio de Janeiro; 1995.*